

Contents lists available at ScienceDirect

### Science of the Total Environment



journal homepage: www.elsevier.com/locate/scitotenv

# Spatial homogeneity and heterogeneity of ambient air pollutants in Tehran



## Sasan Faridi <sup>a,b</sup>, Sadegh Niazi <sup>c</sup>, Fatemeh Yousefian <sup>b</sup>, Faramarz Azimi <sup>d</sup>, Hasan Pasalari <sup>e</sup>, Fatemeh Momeniha <sup>e</sup>, Adel Mokammel <sup>f</sup>, Akbar Gholampour <sup>g</sup>, Mohammad Sadegh Hassanvand <sup>a,b,\*</sup>, Kazem Naddafi <sup>a,b,\*</sup>

<sup>a</sup> Centre for Air Pollution Research (CAPR), Institute for Environmental Research (IER), Tehran University of Medical Sciences, Tehran, Iran

<sup>b</sup> Department of Environmental Health Engineering, School of Public Health, Tehran University of Medical Sciences, Tehran, Iran

<sup>c</sup> International Laboratory for Air Quality and Health, Queensland University of Technology (QUT), Brisbane, Queensland, Australia

<sup>d</sup> Nutrition Health Research Centre, Department of Environment Health, School of Health and Nutrition, Lorestan University of Medical Sciences, Khorramabad, Iran

<sup>e</sup> Department of Environmental Health Engineering, School of Public Health, Iran University of Medical Sciences, Tehran, Iran

<sup>f</sup> Department of Environmental Health Engineering, School of Public Health, Khalkhal University of Medical Sciences, Khalkhal, Iran

<sup>g</sup> Department of Environmental Health Engineering, School of Public Health, Tabriz University of Medical Sciences, Tabriz, Iran

#### HIGHLIGHTS

served in Tehran.

the COD values.

ambient O<sub>3</sub> and PM<sub>10</sub>.

tile, and CV approaches.

A considerable spatial inequality for exposing to ambient air pollutants was ob-

 Ambient PM<sub>10</sub>, PM<sub>2.5</sub> and NO<sub>2</sub> had the highest spatial homogeneity based on

 Regarding CV, the highest and lowest spatial heterogeneity were found for

 There was a remarkable agreement between the results of COD, 90<sup>th</sup> percen-

- GRAPHICAL ABSTRACT
- And Andrew
   Andrew

   Andrew
   Andrew

#### ARTICLE INFO

Article history: Received 10 June 2019 Received in revised form 14 July 2019 Accepted 25 August 2019 Available online 26 August 2019

Editor: Pavlos Kassomenos

Keywords: Spatial variability Homogeneity Heterogeneity

#### ABSTRACT

To investigate spatial inequality of ambient air pollutants and comparison of their heterogeneity and homogeneity across Tehran, the following quantitative indicators were utilized: coefficient of divergence (COD), the 90<sup>th</sup> percentile of the absolute differences between ambient air pollutant concentrations and coefficient of variation (CV). Real-time hourly concentrations of particulate matter (PM) and gaseous air pollutants (GAPs) of twenty-two air quality monitoring stations (AQMSs) were obtained from Tehran Air Quality Control Company (TAQCC) in 2017. Annual mean concentrations of PM<sub>2.5</sub>, PM<sub>10-2.5</sub>, and PM<sub>10</sub> (PM<sub>x</sub>) ranged from 21.7 to 40.5, 37.3 to 75.0 and 58.0 to 110.4  $\mu$ g m<sup>-3</sup>, respectively. Annual mean PM<sub>2.5</sub> and PM<sub>10</sub> concentrations were higher than the World Health Organization air quality guideline (WHO AQG) and national standard levels. NO<sub>2</sub>, O<sub>3</sub>, SO<sub>2</sub> and CO annual mean concentrations ranged from 27.0 to 76.8, 15.5 to 25.1, 4.6 to 12.2 ppb, and 1.9 to 3.8 ppm over AQMSs, respectively. Our generated spatial maps exhibited that ambient PM<sub>x</sub> in Tehran. O<sub>3</sub> hotspots

\* Corresponding authors at: 8th Floor, No. 1547, North Kargar Avenue, Tehran I.R., Iran. E-mail addresses: Hassanvand@tums.ac.ir (M.S. Hassanvand), Knadafi@tums.ac.ir (K. Naddafi). Ambient air pollutants Tehran were observed in the north and south-west, while NO<sub>2</sub> hotspots were in the west and south. COD values of PM<sub>X</sub> demonstrated more results lower than the 0.2 cut off compared to GAPs; indicating high to moderate spatial homogeneity for PM<sub>X</sub> and moderate to high spatial heterogeneity for GAPs. Regarding CV approach, the spatial variabilities of air pollutants followed in the order of O<sub>3</sub> (87.3%) > SO<sub>2</sub> (65.2%) > CO (61.8%) > PM<sub>10-2.5</sub> (52.5%) > PM<sub>2.5</sub> (48.9%) > NO<sub>2</sub> (48.1%) > PM<sub>10</sub> (42.9%), which were mainly in agreement with COD results, except for NO<sub>2</sub>. COD values observed a statistically (*P* < 0.05) positive correlation with the values of the 90<sup>th</sup> percentile across AQMSs. Our study, for the first time, highlights spatial inequality of ambient PM<sub>X</sub> and GAPs in Tehran in detail to better facilitate establishing new intra-urban control policies.

© 2019 Elsevier B.V. All rights reserved.

#### 1. Introduction

As a sobering reality, exposure to ambient air pollution and its health effects has become the most notable environmental risk factor, particularly in developing countries (Huang et al., 2014, Cheng et al., 2016, Marlier et al., 2016, Song et al., 2016, Cohen et al., 2017, Faridi et al., 2018, Jiang and Bai, 2018, Schraufnagel et al., 2019). More recently, the Global Exposure Mortality Model (GEMM) estimated that 8.9 million premature deaths in 2015 were attributed to ambient PM<sub>2.5</sub> (Burnett et al., 2018). Tehran as the capital city of Iran has been receiving the great attention from the Iranian central government, public as well as national and international environmental researchers since it has faced intense ambient air pollution due to unsustainable development, incompatible and/ or non-existent ambient air quality standards along with continued urbanisation, increasing mobile sources and associated emissions (Naddafi et al., 2012; Amini et al., 2017; Faridi et al., 2017; Faridi et al., 2018; Heger and Sarraf, 2018; Hoseini et al., 2018; Mohammadiha et al., 2018). Tehran with >17 million trips and approximately 7 million kilometers traveled per day has specific spatiotemporal traffic conditions. Moreover, daily gasoline, diesel and compressed natural gas consumptions in Tehran were reported approximately 11 and 3 million litres and 2 million cubic meters, respectively (Shahbazi et al., 2016a; Heger and Sarraf, 2018; Mohammadiha et al., 2018; http://traffic.tehran.ir/Default.aspx? tabid=152, 2019). Although numerous studies have been conducted in Tehran regarding various issues of ambient air pollution including investigation of chemical and toxicological characterization of ambient PM (Hassanvand et al., 2014; Hassanvand et al., 2015; Arhami et al., 2018; Faraji et al., 2018; Al Hanai et al., 2019), source apportionment of ambient PM (Taghvaee et al., 2018a, 2018b, 2018c), and ambient air pollutants-related health effects (Naddafi et al., 2012; Faridi et al., 2018; Yousefian et al. 2018), studies that investigate the spatial variation of air pollutants are relatively scarce, to date (Amini et al., 2014, Alizadeh-Choobari et al., 2016, Shahbazi et al., 2018). Exploring and understanding the spatiotemporal variations of ambient air pollutants in intra-urban areas is exceedingly important for controlling and reducing ambient air pollutants' levels, promoting air quality status and protecting public health with the development of appropriate sustainable control strategies and policies, since ambient air pollution is considered as a modifiable and reversible risk factor (Chai et al., 2014; Tan et al., 2014; Zhang and Cao, 2015; Marzouni et al., 2016; Li et al., 2017; Ji et al., 2019). Furthermore, for long- and short-term epidemiological reports as the health impacts studies, it is essential to apply spatiotemporal variabilities of ambient air pollutants (Wang et al., 2014; Dastoorpoor et al., 2016). Three quantitative indicators including COD, the 90<sup>th</sup> percentile of the absolute differences between ambient air pollutant concentrations and CV have extensively been used to better illustrate the spatial variabilities of ambient air pollutants across a city (Pinto et al., 2004; Wilson et al., 2005; Ghim et al., 2015; Tiwari et al., 2015; Guo et al., 2017; He et al., 2017; Jin et al., 2017; Song et al., 2017; Zhao et al., 2018; Saha et al., 2019). COD as an indicator has extensively used in numerical analyses to determine the resemblance between two different under study datasets (Kim et al., 2005). In fact, the values of COD provide more information regarding the degree of uniformity between the concentrations of ambient air pollutants, measuring simultaneously at various AQMSs across a city (Wongphatarakul et al., 1998; Kim et al., 2005; Krudysz et al., 2008). Besides, the 90<sup>th</sup> percentile of the absolute differences between ambient air pollutant concentrations across AQMSs was utilized to show an absolute measure of the degree of heterogeneity and homogeneity in their concentrations within a city (Pinto et al., 2004; Ghim et al., 2015). Moreover, CV indicator describes the dispersion degree of ambient air pollutant data across AQMSs (He et al., 2017; Song et al., 2017). However, it should be emphasized that using each approach alone might not be sufficient to determine the ambient air pollutant's homogeneity and heterogeneity across urban areas (Pinto et al., 2004). Consequently, investigation of spatial characteristics of ambient air pollutants using three popular methods mentioned above is therefore essential to create better awareness of the air quality status in different districts of urban areas to inform authorities and policy-makers for vigorously implementing the governmental and societal strategies of joint prevention and control of ambient air pollution and its health consequences.

Based on the above-mentioned concerns, the major objectives of this study are to **1**) investigate the spatial variations and depict hotspots of ambient  $PM_X$  and GAPs in Tehran in 2017, and **2**) compare their heterogeneity and homogeneity across different districts of Tehran based on COD, the 90<sup>th</sup> percentile of the absolute differences between ambient air pollutant concentrations and CV approaches across validated AQMSs.

#### 2. Materials and methods

#### 2.1. Air quality data gathering and processing

Real-time hourly average concentrations of each criteria air pollutant (PM<sub>2.5</sub>, PM<sub>10</sub>, NO<sub>2</sub>, O<sub>3</sub>, SO<sub>2</sub>, and CO) for the year 2017 were obtained from the twenty-two ground-based AQMSs (Table 1) which belong to TAQCC (Shahbazi et al., 2018, http://airnow.tehran.ir/home/ dataarchive.aspx, 2019). Fig. S1 (from Supplementary Information) shows the spatial distribution of all AQMSs in different districts of Tehran. At AQMSs, PM<sub>2.5</sub> and PM<sub>10</sub>, O<sub>3</sub>, NO<sub>2</sub>, SO<sub>2</sub>, and CO are monitored using the beta-attenuation, UV-spectrophotometry, chemiluminescence, ultraviolet fluorescence, and non-dispersive infrared absorption methods, respectively. As shown in Table 1, the hourly data coverage in active AQMSs ranged from 66.6 to 97.6%, 46.3 to 97.7%, 52.5 to 99.8%, 17.1 to 99.9%, 18.3 to 98.5%, and 36.8 to 98.5% in 2017 for PM<sub>2.5</sub>,  $PM_{10}$ ,  $O_3$ ,  $NO_2$ ,  $SO_2$ , and CO, respectively. In the current study, air quality data processing at AQMSs has been checked based on the z-score method to reject spatial and temporal outliers (He et al., 2017; Song et al., 2017; Faridi et al., 2018; van Zoest et al., 2018). It should be highlighted that aforementioned air quality data processing method was only used on AQMSs with >75% completeness of the total hours

| Та | ble | 1 |
|----|-----|---|
|    |     |   |

Detailed information on AQMSs and their hourly data coverage for each ambient air pollutant.

| No. | AQMSs           | Location |           | Hourly data coverage (%) <sup>a</sup> |                  |             |                 |                 |             |
|-----|-----------------|----------|-----------|---------------------------------------|------------------|-------------|-----------------|-----------------|-------------|
|     | Name            | Latitude | Longitude | PM <sub>2.5</sub>                     | PM <sub>10</sub> | 03          | NO <sub>2</sub> | SO <sub>2</sub> | CO          |
| 1   | Roz Park        | 35.739   | 51.267    | 8305 (94.8)                           | 8478 (96.8)      | 6779 (77.4) | 2010 (22.9)     | 7478 (85.4)     | 8442 (96.4) |
| 2   | Poonak          | 35.762   | 51.331    | 7264 (82.9)                           | 5841 (66.7)      | 5484 (62.6) | 7509 (85.7)     | 6206 (70.8)     | 7209 (82.3) |
| 3   | Zone 2          | 35.777   | 51.368    | 7980 (91.1)                           | 6931 (79.1)      | 6362 (72.6) | 1494 (17.1)     | 1605 (18.3)     | 8177 (93.3) |
| 4   | Darous          | 35.770   | 51.454    | 0                                     | 0                | 0           | 0               | 0               | 0           |
| 5   | Golbarg         | 35.731   | 51.506    | 7960 (90.9)                           | 8072 (92.1)      | 7782 (88.8) | 8009 (91.4)     | 5864 (66.9)     | 8455 (96.5) |
| 6   | Setad bohran    | 35.727   | 51.431    | 8506 (97.1)                           | 7945 (90.7)      | 7604 (86.8) | 7562 (86.3)     | 7229 (82.5)     | 7875 (89.9) |
| 7   | Sharif          | 35.702   | 51.350    | 8497 (97.0)                           | 8559 (97.7)      | 8740 (99.8) | 8260 (94.3)     | 5700 (65.1)     | 5873 (67.0) |
| 8   | Tarbiat Modares | 35.717   | 51.385    | 8460 (96.6)                           | 8503 (97.1)      | 5401 (61.7) | 8660 (98.9)     | 8628 (98.5)     | 8550 (97.6) |
| 9   | Piroozi         | 35.695   | 51.493    | 7227 (82.5)                           | 8486 (96.9)      | 7388 (84.3) | 8489 (96.9)     | 8291 (94.6)     | 8569 (97.8) |
| 10  | Fath            | 35.678   | 51.337    | 0                                     | 7659 (87.4)      | 7703 (87.9) | 7830 (89.4)     | 4896 (55.9)     | 5725 (65.4) |
| 11  | Zone 11         | 35.672   | 51.389    | 6723 (76.7)                           | 0                | 0           | 4008 (45.8)     | 7010 (80.0)     | 8156 (93.1) |
| 12  | Zone 16         | 35.644   | 51.397    | 0                                     | 0                | 5287 (60.4) | 8384 (95.7)     | 6322 (72.2)     | 8475 (96.7) |
| 13  | Shad abad       | 35.670   | 51.297    | 8142 (92.9)                           | 8549 (97.6)      | 8504 (97.1) | 8757 (99.9)     | 7985 (91.2)     | 0           |
| 14  | Zone 19         | 35.635   | 51.362    | 0                                     | 6884 (78.6)      | 0           | 0               | 6662 (76.1)     | 4106 (46.9) |
| 15  | Masoudiyeh      | 35.630   | 51.499    | 8225 (93.9)                           | 8050 (91.9)      | 5680 (64.8) | 6035 (68.9)     | 7696 (87.9)     | 8554 (97.6) |
| 16  | Share rey       | 35.603   | 51.425    | 8486 (96.9)                           | 7771 (88.7)      | 7697 (87.9) | 6979 (79.7)     | 7967 (90.9)     | 8506 (97.1) |
| 17  | Zone 4          | 35.741   | 51.506    | 5832 (66.6)                           | 0                | 5270 (60.2) | 7072 (80.7)     | 6364 (72.6)     | 7184 (82.0) |
| 18  | Zone 10         | 35.697   | 51.358    | 7971 (91.0)                           | 0                | 0           | 0               | 8186 (93.4)     | 0           |
| 19  | Mahalati        | 35.661   | 51.466    | 0                                     | 4059 (46.3)      | 0           | 3487 (39.8)     | 4489 (51.2)     | 4270 (48.7) |
| 20  | Tehransar       | 35.697   | 51.243    | 8008 (91.4)                           | 8077 (92.2)      | 4598 (52.5) | 8300 (94.7)     | 8091 (92.4)     | 3220 (36.8) |
| 21  | Aghdasiyeh      | 35.795   | 51.484    | 8056 (92.0)                           | 8225 (93.9)      | 0           | 4011 (45.8)     | 0               | 8625 (98.5) |
| 22  | Sadr            | 35.778   | 51.428    | 8548 (97.6)                           | 0                | 5966 (68.1) | 8109 (92.6)     | 5094 (58.2)     | 7507 (85.7) |

<sup>a</sup> AQMSs with <75% valid hourly data available were excluded from the next analysis.

in 2017 for ambient PM<sub>X</sub> and GAPs (De Hoogh et al., 2018). As can be seen from Table 1, based on the above-mentioned inclusion criterion, there were 16, 14, 9, 13, 13 and 14 eligible stations for ambient PM<sub>2.5</sub>. PM<sub>10</sub>, O<sub>3</sub>, NO<sub>2</sub>, SO<sub>2</sub>, and CO, respectively. Furthermore, the percentage of the missing data for ambient PM<sub>X</sub> and GAPs in Tehran measured by the eligible stations was mainly <15% of total hours in 2017. Herein, the method of air quality data control is similar to that of our previous study (Faridi et al., 2018), with the exception of the second stage in zscore method. Firstly, the time series of hourly concentrations of each air pollutant was standardized using the z-score method, and then the air quality data were used in the subsequent computations if they meet the following conditions: **1**) an absolute z-score  $<4(|Z_t| < 4)$ ; **2**) the increment from the previous hourly value being  $< 6 (Z_t - Z_{t-1} < 6)$ ; and 3) and the ratio of the z-score to its centered rolling average of order 3 (RA3) <2 (Z<sub>t</sub> / RA3(Z<sub>t</sub>) < 2) (He et al., 2017; Song et al., 2017). Consequently, <0.3% of air quality data for each air pollutant was removed from the hourly datasets and the subsequent computations. As such, after air quality data processing and cleaning, all eligible AQMSs were accepted as the validated AQMSs, and an annual average concentration was calculated for each mentioned air pollutant at station-level. Additionally, the hourly average concentration of each ambient air pollutant at a total city-level was calculated using the average of all hourly concentrations across all validated AQMSs. For the first time, herein, we reported the concentration of coarse particles; particles with an aerodynamic diameter between 2.5 and 10  $\mu$ m (PM<sub>10-2.5</sub>); at AQMSs and citylevel in Tehran for the year 2017. PM<sub>10-2.5</sub> concentrations were calculated by the difference between the concentrations of PM<sub>10</sub> and PM<sub>2.5</sub> at 12 validated AQMSs (Hassanvand et al., 2017; Wang et al., 2018).

#### 2.2. Quantifying spatial variability

To examine spatial variabilities of ambient air pollutants and their heterogeneity and homogeneity in different districts of Tehran, the three quantitative indicators were used: 1) COD (Amini et al., 2017; Song et al., 2017); 2) 90<sup>th</sup> percentile of the absolute differences between 24-h (PM<sub>X</sub>, NO<sub>2</sub>, SO<sub>2</sub>) and 8-h (O<sub>3</sub> and CO) air pollutant concentrations across AQMSs for a given year (Pinto et al., 2004; Mangia et al., 2013; Ghim et al., 2015); and 3) CV (He et al., 2017; Jin et al., 2017; Song et al., 2017). COD values were computed from validated AQMSs as

follows (Wongphatarakul et al., 1998; Kim et al., 2005; Krudysz et al., 2008):

$$COD_{jk} = \sqrt{\frac{1}{p}\sum_{i=1}^{p}\left(\frac{APC_{ij} - APC_{ik}}{APC_{ij} + APC_{ik}}\right)^2}$$

Where, APC<sub>ii</sub> and APC<sub>ik</sub> represent ambient air pollutant concentrations for time i (i for PM<sub>x</sub>, NO<sub>2</sub>, and SO<sub>2</sub> is 24-h average concentrations, for  $O_3$  and CO is moving 8-h average concentrations) at station j and k, j and k represent two datasets for two given AQMSs, and p is the number of 24- and 8-h average concentrations of ambient air pollutants for the whole year. If the two datasets related to air pollutant concentrations are similar, COD value approaches zero, otherwise, COD approaches one. Further, exactly similar to the complex network correlation model developed by N.-N. Zhang et al. (2019) and N.-N. Zhang et al. (2018) (Zhang et al., 2018; Zhang et al., 2019), we developed and created a complex network COD model to show heterogeneity and homogeneity of all ambient air pollutants across validated AQMSs in Tehran for 2017 using the igraph R Package. The following equation was used to calculate the 90th percentile of the absolute differences between ambient air pollutant concentrations at AQMSs (Pinto et al., 2004; Ghim et al., 2015):

 $90^{th} percentile = ABS(APC_{ij} - APC_{ik})$ 

 $APC_{ij}$  and  $APC_{ik}$  have previously been defined in detail. The CV indicator was calculated using the following equation (He et al., 2017; Song et al., 2017):

$$CV = \frac{SID}{\overline{X}}$$
$$STD = \sqrt{\frac{1}{p} \sum_{i=1}^{p} (APC_i - \overline{X})^2}$$

Where, STD is the standard deviation, APC<sub>i</sub> represents ambient air pollutant concentrations for time *i* at AQMSs,  $\overline{X}$  represents annual



mean concentration for each air pollutant. In this study, to better show the correlation between the values of COD and 90<sup>th</sup> percentile, we introduced a Pearson correlation between them as a novel approach (Wang et al., 2014; Zhao et al., 2016; Xu et al., 2018). *P*-values <0.05 were considered as statistically significant.

#### 3. Results and discussion

#### 3.1. Spatial variabilities and hotspots of ambient air pollutants

Fig. 1 (a to g) reveals the annual average data on ambient  $PM_x$  as well as on GAPs for all validated ground-based AQMSs in Tehran for 2017. In addition, to better highlight the most polluted areas and the changes in annual mean concentrations of ambient air pollutants, the map of spatial distribution of annual mean concentrations of ambient PM<sub>x</sub> and GAPs was generated by using a Geographic Information System (Fig. 2). As shown in Fig. 1a, annual mean PM<sub>2.5</sub> concentrations measured at all AQMSs ranged from 21.7 to 40.5  $\mu g~m^{-3}$  , and are all approximately 2–4 times higher than the WHO AQG ( $10 \,\mu g \, m^{-3}$ ), U.S. EPA and national levels ( $12 \,\mu g \, m^{-3}$ ). Unfortunately, even the  $25^{th}$  percentile of mean PM<sub>2.5</sub> concentrations at all AQMSs exceeded above-mentioned recommendation and standards (Fig. 1a and Table S1). Interestingly, Tehran citizens who live around Shad Abad (40.5  $\mu$ g m<sup>-3</sup>) and Zone 11 (40.3  $\mu$ g m<sup>-3</sup>) stations were exposed to PM<sub>2.5</sub> at approximately two times higher levels compared with those in Zone 2 (21.7  $\mu$ g m<sup>-3</sup>) and Golbarg (23.6  $\mu$ g m<sup>-3</sup>) stations (Fig. 2). On the other hand, the highest to lowest annual mean ratio across AQMSs was around 2 and it is a clear inequality regarding exposure to ambient PM<sub>2.5</sub>. Annual mean concentrations of PM<sub>2.5</sub> at 8 stations (out of 16 stations) including Setad Bohran, Zone 10, Poonak, Roz Park, Aghdasiyeh, Masoudiyeh, Golbarg and Zone 2 were less than the overall mean concentration  $(32.1 \ \mu g \ m^{-3})$  for Tehran, whereas annual mean PM<sub>2.5</sub> concentrations at the other AQMSs were more than the overall mean concentration (Fig. 1a and 2). As illustrated in Fig. 2, PM<sub>2.5</sub> showed the highest annual mean concentrations in the south, south-east and central AQMSs. Further, annual average concentrations of PM<sub>2.5</sub> rose from the northern areas of Tehran megacity to its south (Fig. 2). Annual mean concentrations of PM<sub>10</sub> spanned from 58.0 to 110.4  $\mu$ g m<sup>-3</sup> at all AQMSs, thereby 2.9–5.5 times higher than the WHO AQG (20  $\mu$ g m<sup>-3</sup>) (Fig. 1b and Table S2). Similar to PM<sub>2.5</sub>, the 25<sup>th</sup> percentile of annual mean PM<sub>10</sub> concentrations observed at all AQMSs was more than the WHO recommendation. Annual mean PM<sub>10</sub> concentrations observed at Fath, Piroozi, Shad Abad, Shahr Rey, Tehransar, Sharif and Tarbiat Modares stations were greater than that of the overall mean concentration (84.9  $\mu g m^{-3}$ ) surveyed in Tehran in 2017 (Fig. 1a and 2). On the other hand, annual mean PM<sub>10</sub> concentrations increased from the central to south areas, with the maximum in Share Rev station with 110.4  $\mu g m^{-3}$ , followed by Tehransar station with 110.1  $\mu g m^{-3}$  in the eastern areas of Tehran, as illustrated in Fig. 2. Similar to PM<sub>2.5</sub>, exposure to PM<sub>10</sub> indicated a clear spatial inequality across Tehran with the ratio of highest to lowest annual mean >2 (Fig. 2 and Table S2). In terms of PM<sub>10-2.5</sub>, annual mean concentrations of PM<sub>10-2.5</sub> varied from 37.3  $\mu g m^{-3}$  in Setad bohran station to 75.0  $\mu g m^{-3}$  in Tehransar (Fig. 1c). As shown in Fig. 2 and Table S3, the value of inequality exposure to  $PM_{10-2.5}$  between stations with the highest and lowest  $PM_{10-2.5}$  annual mean in Tehran was nearly 2. Alike to PM<sub>10</sub>, the highest annual mean concentrations of ambient PM<sub>10-2.5</sub> were found in the eastern and southern areas of Tehran megacity with 75.0 and 73.1  $\mu$ g m<sup>-3</sup> for Tehransar and Shahr Rey stations, respectively (Fig. 1c). The annual mean NO<sub>2</sub> concentrations exceeded the WHO AQG (22 ppb), by approximately 1.2- (Shad Abad with 27.0 ppb) to 3.5- (Zone 16 with 76.8 ppb) times (Fig. 1d and Table S4). Compared with the U.S. EPA and national levels (53 ppb), annual mean NO<sub>2</sub> concentration at 6 AQMSs (out of 13) in Tehran was less than the above-mentioned standard levels in 2017, whereas annual mean NO<sub>2</sub> concentration measured at 7 stations was more than the standard levels. Annual mean concentrations of O<sub>3</sub> were in the range of 15.5 (Shahr Rey station) -25.1 ppb (Piroozi station) (Fig. 1e and Table S5). As expected, O<sub>3</sub>-8h concentrations displayed a different pattern of distribution to other air pollutants throughout Tehran (Fig. 2). O<sub>3</sub> concentrations can be decreased by high NO<sub>x</sub> emissions as a result of titration effects (Jhun et al., 2015). Zone 16 and Zone 4 showed the aforementioned effect, with high concentrations in NO<sub>2</sub>. but low  $O_3$  (Fig. 2). The highest  $O_3$  concentrations were observed in the northern and south-western parts of Tehran megacity; while the central areas demonstrated low to moderate concentrations (Fig. 2). NO<sub>2</sub> concentrations were higher in the west and south, while the eastern stations displayed the lowest concentrations (Fig. 2). In relation to stations-specific concentrations of SO2, annual mean concentrations of SO<sub>2</sub> were between 4.6 ppb in Roz Park station and 12.2 ppb in Tehransar station (Fig. 1f and Table S6), while the CO concentrations at all AOMSs ranged from approximately 1.9 ppm (Roz Park) to 3.8 ppm (Zone 16) (Fig. 1g). Most areas of the city exhibited an acceptable concentration for SO<sub>2</sub> except for Tehransar (12.2 ppb) and Zone 16 (11.4 ppb), in the east and centre, respectively; these monitoring stations recorded the highest concentrations (Fig. 2). Such high figures can be attributed to the industry located within these regions of Tehran (Seifi et al., 2019). The detailed descriptive statistics are also provided in Table S1 to S7.

#### 3.2. Spatial homogeneity and heterogeneity of ambient air pollutants

Fig. 3 (a to g) illustrates the network diagram of homogeneous and heterogeneous AQMSs based on COD values with 0.2 cut off for ambient air pollutants across Tehran. In addition, the values of COD and 90<sup>th</sup> percentile are shown in the Table S8 to S14 in detail. As discussed earlier, COD is an important factor to see the differences in absolute concentrations or source contribution, as a high r value does not necessarily indicate homogeneity of concentrations (Cyrys et al., 2008). Consequently, intra-urban regions with low absolute differences in ambient air pollutant concentrations tended to experience low COD values; <0.2. COD values of PM<sub>X</sub> demonstrated more results lower than the 0.2 cut off compared to GAPs, indicating spatial homogeneity in the concentrations of ambient PM<sub>X</sub> across the validated AQMSs in Tehran (Table S8 to S14). Additionally, the boxplots of COD values for both PM<sub>X</sub> and GAPs are shown in Fig. S2. Overall, PM<sub>10</sub> had the lowest mean COD value with 0.181, whereas  $SO_2$  had the highest mean COD (0.318). For  $PM_{2.5}$ , PM<sub>10-2.5</sub>, NO<sub>2</sub>, O<sub>3</sub> and CO, mean COD was overall 0.216, 0.240, 0.292, 0.287 and 0.293, respectively (Table S15). Also, median COD values were above 0.2 for all ambient air pollutants, except for PM<sub>10</sub>. However, even with twenty-two AQMSs across 730 km<sup>2</sup> of Tehran, COD values among some AQMSs indicate a considerable heterogeneous spatial distribution of air pollutants in different districts of Tehran. COD values ranged from 0.106 to 0.379 for  $PM_{2.5}$ , 0.068 to 0.328 for  $PM_{10}$  and 0.139 to 0.482 for PM<sub>10-2.5</sub> (Table S8 to S10). On the subject of PM<sub>2.5</sub>, Setad Bohran and Sadr stations showed the highest homogeneity and heterogeneity with other stations, at 12 and 15 stations, respectively (Fig. 3a and Table S8). For  $PM_{10}$ , there was approximately a balance in homogeneity and heterogeneity, as Roz Park, Piroozi, Sharif and Tarbiat Modares showed the maximum homogeneity with 10 stations. With regard to PM<sub>10</sub>, Aghdasiyeh station with 8 out of 13 stations showed the highest number of heterogeneity (Fig. 3b and Table S9). The number of heterogeneous stations for PM<sub>10-2.5</sub> was higher compared to other PM fractions; PM<sub>2.5</sub> and PM<sub>10</sub>. Regarding PM<sub>10-2.5</sub>, Sharif, Tarbiat Modares and Roz Park were spatially homogenous with 6 stations (Fig. 3c and Table S10). According to Table S8, for 24-h PM<sub>25</sub>

Fig. 1. The station-specific boxplot of air pollutant concentrations (PM<sub>2.5</sub> (a), PM<sub>10</sub> (b), PM<sub>10-2.5</sub> (c), NO<sub>2</sub> (d), O<sub>3</sub> (e), SO<sub>2</sub> (f), CO (g)) in Tehran (2017). Multiplication sign (X) in each box is the annual mean concentration of air pollutant at each AQMS.



Fig. 2. Spatial distribution of annual mean concentration of ambient air pollutants across Tehran (2017).

Fig. 3. Network diagram of homogeneous and heterogeneous AQMSs based on COD values with 0.2 cut off for ambient air pollutants (PM<sub>2.5</sub> (a), PM<sub>10</sub> (b), PM<sub>10-2.5</sub> (c), NO<sub>2</sub> (d), O<sub>3</sub> (e), SO<sub>2</sub> (f), CO (g)) across validated ground-based AQMSs within Tehran megacity (2017).



concentrations, Tarbiat Modares and Shahr Rey stations reveal the strong similarity between the two stations with a COD value equal to 0.106, followed by Tarbiat Modares and Sharif stations with COD value equal to 0.109. Furthermore, of all AQMSs across Tehran examined, the two most dissimilar monitoring stations were Zone 2 and Sadr (COD = 0.379), as well as Zone 2 and Zone 11 (0.357). Regarding 24h PM<sub>10</sub> concentrations, as shown in Table S9, among all studied AQMSs, Shad Abad and Fath stations exhibited the greatest similarity with a COD value of 0.068, followed by Shad Abad and Tehransar, Tarbiat Modares and Tehransar, Tarbiat Modares and Sharif stations with a COD values of 0.076, 0.078, and 0.078, respectively. Of all paired AQMSs investigated throughout Tehran, the highest spatial variability of 24-h PM<sub>10</sub> concentrations was recorded between Zone 2 and Shar Rey stations with COD of 0.328, as well as Zone 2 and Tehransar stations with COD equal to 0.310 (Table S9). In regards to 24-h  $PM_{10-2.5}$ , the highest spatial uniformity in PM<sub>10-2.5</sub> concentrations was observed between Piroozi and Golbarg stations (COD = 0.139), followed by Tehransar and Shad Abad stations with a COD value equal to 0.143. In addition, the highest COD (0.482) for PM<sub>10-2.5</sub> concentrations was calculated for Piroozi and Shad Abad stations (Table S10). There are a number of reasons that could account for spatial variation in PM<sub>x</sub> concentrations. Sporadic events such as land uses, construction and demolition activities, proximity to emission sources, transient emission events, or a meteorological occurrence, such as local circulations and topographic features, are the most significant reasons (Pinto et al., 2004). Ambient PM<sub>x</sub> comprises a mixture of primary and secondary constituents. Significant differences on spatial and temporal characteristics can also be attributed to local sources of ambient PM<sub>X</sub>. Moreover, based on recent conducted study by Taghvaee et al. (2018a, 2018b, 2018c), the source of ambient PM<sub>2.5</sub> in two central areas of Tehran arises mostly from mobile sources; an overall contribution equal to 77% and 73%, a high contribution of 74% and 63% in the warm season and 78% and 80% in the cold season (Taghvaee et al., 2018a, 2018b, 2018c). Also, previously conducted studies have reported these high contributions of vehicular emissions to ambient PM air pollutants (Shahbazi et al., 2016b, Arhami et al., 2017, Arhami et al., 2018, Taghvaee et al., 2018a, 2018b, 2018c). Therefore, high COD values suggest that the available groundbased AQMSs may be insufficient to accurately monitor of the whole area of the city (Guo et al., 2017). Compared to ambient PM<sub>x</sub>, GAPs showed higher spatial variabilities between paired intra-urban AQMSs over districts of Tehran (Fig. 3d to g). In reality, CODs of GAPs revealed a higher heterogeneous spatial distribution between monitoring stations. As indicated in Table S11 to 14, NO<sub>2</sub> with more figures <0.2 had the highest spatial uniformity in comparison to other ambient GAPs. For NO<sub>2</sub>, the maximum homogeneity was observed between Setad Bohran station and other five stations including Zone 4 (0.138), Tehransar (0.142), Sharif (0.150), Zone 16 (0.158), and Shahr Rey (0.182). The highest COD value calculated for Shad Abad and Zone 4 stations; 0.579 (Table S11). Moreover, Golbarg, Piroozi and Shad Abad stations were spatially heterogeneous with others (Fig. 3d). Similar to  $O_3$ , 24-h SO<sub>2</sub> and 8-h CO represented COD values >0.2 which might be associated with specific source emissions. Regarding 8-h O<sub>3</sub> (Fig. 3e to g and Tables S12 to S14), Shad Abad exhibited a low COD value, with Fath and Sharif exhibiting 0.194 and 0.181, respectively. Moreover, we observed a considerable homogeneity between Fath and Sharif stations with a COD value equal to 0.180. The results of COD for 24-h SO<sub>2</sub> concentrations demonstrated that only 5 out of 13 stations, such as Setad Bohran, Shad Abad, Shahr Rey, Zone 4 and Tarbiat Modares were spatially homogenous (Fig. 3f). 8-h CO concentrations displayed a high spatial heterogeneity at 8 stations in comparison with others (Fig. 3g). Further details about heterogeneous and homogenous sites are available in Table S8 to S14. In addition to calculation of COD for ambient PM and GAPs, herein we computed the 90<sup>th</sup> percentile of the absolute differences in air pollutant concentrations intra-urban of Tehran to show their spatial uniformity and dissimilarity. The results of the 90<sup>th</sup> percentile of the absolute differences in air pollutant concentrations across

#### Table 2

Pearson correlation coefficients between COD and  $90^{\text{th}}$  percentile in Tehran (2017), \* *P* < 0.05.

| COD 90 <sup>th</sup><br>percentile   | PM <sub>2.5</sub> | PM <sub>10</sub> | PM <sub>10-2.5</sub> | 03      | NO <sub>2</sub> | SO <sub>2</sub> | СО      |
|--|-------------------|------------------|----------------------|---------|-----------------|-----------------|---------|
| $\begin{array}{c} PM_{2.5} \\ PM_{10} \\ PM_{10-2.5} \\ O_3 \\ NO_2 \\ SO_2 \\ CO \end{array}$ | 0.8905*           | 0.8361*          | 0.6095*              | 0.4874* | 0.4840*         | 0.7894*         | 0.8918* |

AQMSs are shown in Table S8 to S14. Our results revealed that when COD values decline, the values of 90th percentile of the absolute differences decreased, or vice versa (Table 2). Consequently, we introduced Pearson correlation between the values of COD and the 90<sup>th</sup> percentile of the absolute differences between ambient air pollutant concentrations as a novel approach to investigate the correlation of these approaches. Interestingly, there was a statistically positive correlation (P < 0.05) between the values of COD and the 90<sup>th</sup> percentile, indicating agreement between these approaches, as illustrated in Table 2. The coefficients of Pearson correlation between the values of COD and the 90<sup>th</sup> percentile for  $PM_{2.5}$  (0.8905) and  $PM_{10}$  (0.8361), as the most notable markers of ambient air pollution globally, were higher than those for GAPs, including O<sub>3</sub>, NO<sub>2</sub> and SO<sub>2</sub> (Table 2). Furthermore, similar to PM<sub>2.5</sub> and PM<sub>10</sub>, CO was found with high coefficients (0.8918) of Pearson correlation between the values of COD and the 90<sup>th</sup> percentile. These findings confirm that in order to better understanding spatial variability of all ambient air pollutants, in addition to COD and the 90<sup>th</sup> percentile, the investigation of correlation between them could be appropriate. Fig. S3 summarizes the boxplot of the 90<sup>th</sup> percentile of the absolute differences between 24-h concentrations for  $PM_X$ ,  $SO_2$ , NO<sub>2</sub> and 8-h concentrations for O<sub>3</sub> and CO. As shown in Fig. S3 and Table S16, the values of the 90<sup>th</sup> percentile of the absolute differences between 24-h concentrations for PM<sub>2.5</sub>, PM<sub>10</sub> and PM<sub>10-2.5</sub> was approximately in the range of 10.6–39.4, 15.4–100.1, and 18.1–79.9  $\mu$ g m<sup>-3</sup>, with the mean value of approximately 23.3, 49.7 and 39.3  $\mu$ g m<sup>-3</sup>, respectively. Furthermore, the mean value of the 90<sup>th</sup> percentile of the absolute differences for O<sub>3</sub>, NO<sub>2</sub>, SO<sub>2</sub> and CO was approximately 23.2, 46.9 10.0 ppb and 2.6 ppm, respectively. Detailed information regarding the 90<sup>th</sup> percentile is presented in Table S16. Based on the CV indicator (Fig. 4), the highest spatial variability at all validated AQMSs was recognized for ambient  $O_3$  with 87.3%, followed by  $SO_2$  (65.2%) > CO (61.8%)  $> PM_{10-25} (52.5\%) > PM_{25} (48.9\%) > NO_2 (48.1\%) > PM_{10} (42.9\%)$ . Interestingly, we observed that among ambient GAPs,  $O_3$  as a secondary air



Fig. 4. The CV of ambient PM<sub>X</sub> and gaseous air pollutants in Tehran (2017).

pollutant had the highest heterogeneity whereas its most important precursor and destroyer; ambient NO<sub>2</sub>; showed the highest homogeneity, or vice versa. The results of spatial distribution from the CV indictor tended to be in agreement with those from COD indicator, except for NO<sub>2</sub>.

#### 4. Conclusion

To better understanding the status of ambient air pollutants in different districts of Tehran, the spatial variabilities of ambient PM<sub>X</sub> and GAPs was investigated with using the approaches of COD, CV and the 90<sup>th</sup> percentile of the absolute differences between air pollutant concentrations across AQMSs. Based on our results, the annual mean concentrations of PM<sub>2.5</sub>, PM<sub>10</sub> and PM<sub>10-2.5</sub>, NO<sub>2</sub>, O<sub>3</sub> and SO<sub>2</sub>, and CO were 32.1, 84.9 and  $53.5 \,\mu g \, m^{-3}$ , 20.7, 53.4 and  $8.1 \, ppb$  and  $2.6 \, ppm$  in Tehran for 2017. Our spatial results demonstrated that Tehran citizens were exposed to various concentrations of ambient PM<sub>x</sub> and GAPs across the different districts of Tehran in 2017. Regarding to annual PM<sub>2.5</sub>, throughout monitoring stations, Shad Abad and Zone 2 observed the highest and lowest concentrations with 40.5 and 21.7  $\mu$ g m<sup>-3</sup>, respectively. Considering PM<sub>10</sub>, annual mean concentrations were in the range of 58.0 (Zone 2) -110.4  $\mu$ g m<sup>-3</sup> (Shahr Rey). Furthermore, Tehran citizens experienced annual mean PM2.5 and PM10 concentrations more than the WHO AQG and national standard levels at all investigated municipal districts of Tehran in 2017. Overall, hotspots for ambient PM<sub>x</sub> pollution were mainly concentrated in the south. south-east and central regions of Tehran city. Considering three quantitative indicators to show spatial inequality of ambient air pollutants, the highest homogeneity was found for ambient PM<sub>x</sub> air pollution in Tehran, whereas the highest heterogeneity was observed for ambient GAPs, except for NO<sub>2</sub> Moreover, the highest homogeneity was found for PM<sub>2.5</sub> and NO<sub>2</sub>, as the most notable markers of traffic-related emissions in different districts of Tehran in 2017. Interestingly, the results of COD revealed that spatial pattern of O<sub>3</sub> and its most important precursor and destroyer; ambient NO<sub>2</sub>; were inverse. The results of this study will help bring awareness to government, policy-makers, medical professionals, and the public about the spatial variation of ambient air pollutants, particularly ambient PM<sub>2.5</sub> as a major health concern globally, to facilitate establishing new local and intra-urban control policies.

#### Acknowledgements

This study was supported by Tehran University of Medical Sciences (grant number 97-03-66-40508). Also, the authors are grateful to Tehran Air Quality Control Company (TAQCC) for providing ambient air pollutants concentrations data.

#### **Declaration of competing interest**

There is no actual or potential conflict of interest among authors.

#### Appendix A. Supplementary data

Supplementary data to this article can be found online at https://doi. org/10.1016/j.scitotenv.2019.134123.

#### References

- Al Hanai, A.H., Antkiewicz, D.S., Hemming, J.D., Shafer, M.M., Lai, A.M., Arhami, M., Hosseini, V., Schauer, J.J., 2019. Seasonal variations in the oxidative stress and inflammatory potential of PM2. 5 in Tehran using an alveolar macrophage model; the role of chemical composition and sources. Environ. Int. 123, 417–427.
- Alizadeh-Choobari, O., Bidokhti, A., Ghafarian, P., Najafi, M., 2016. Temporal and spatial variations of particulate matter and gaseous pollutants in the urban area of Tehran. Atmos. Environ. 141, 443–453.
- Amini, H., Taghavi-Shahri, S.M., Henderson, S.B., Naddafi, K., Nabizadeh, R., Yunesian, M., 2014. Land use regression models to estimate the annual and seasonal spatial vari-

ability of sulfur dioxide and particulate matter in Tehran, Iran. Sci. Total Environ. 488, 343–353.

- Amini, H., Hosseini, V., Schindler, C., Hassankhany, H., Yunesian, M., Henderson, S.B., Künzli, N., 2017. Spatiotemporal description of BTEX volatile organic compounds in a Middle Eastern megacity: Tehran study of exposure prediction for environmental health research (Tehran SEPEHR). Environ. Pollut. 226, 219–229.
- Arhami, M., Hosseini, V., Shahne, M.Z., Bigdeli, M., Lai, A., Schauer, J.J., 2017. Seasonal trends, chemical speciation and source apportionment of fine PM in Tehran. Atmos. Environ. 153, 70–82.
- Arhami, M., Shahne, M.Z., Hosseini, V., Haghighat, N.R., Lai, A.M., Schauer, J.J., 2018. Seasonal trends in the composition and sources of PM 2.5 and carbonaceous aerosol in Tehran, Iran. Environ. Pollut. 239, 69–81.
- Burnett, R., Chen, H., Szyszkowicz, M., Fann, N., Hubbell, B., Pope, C.A., Apte, J.S., Brauer, M., Cohen, A., Weichenthal, S., 2018. Global estimates of mortality associated with longterm exposure to outdoor fine particulate matter. Proc. Natl. Acad. Sci. 115 (38), 9592–9597.
- Chai, F., Gao, J., Chen, Z., Wang, S., Zhang, Y., Zhang, J., Zhang, H., Yun, Y., Ren, C., 2014. Spatial and temporal variation of particulate matter and gaseous pollutants in 26 cities in China. J. Environ. Sci. 26 (1), 75–82.
- Cheng, Z., Luo, L., Wang, S., Wang, Y., Sharma, S., Shimadera, H., Wang, X., Bressi, M., de Miranda, R.M., Jiang, J., 2016. Status and characteristics of ambient PM2. 5 pollution in global megacities. Environ. Int. 89, 212–221.
- Cohen, A.J., Brauer, M., Burnett, R., Anderson, H.R., Frostad, J., Estep, K., Balakrishnan, K., Brunekreef, B., Dandona, L., Dandona, R., 2017. Estimates and 25-year trends of the global burden of disease attributable to ambient air pollution: an analysis of data from the global burden of diseases study 2015. Lancet 389 (10082), 1907–1918.
- Cyrys, J., Pitz, M., Heinrich, J., Wichmann, H.-E., Peters, A., 2008. Spatial and temporal variation of particle number concentration in Augsburg, Germany. Sci. Total Environ. 401 (1–3), 168–175.
- Dastoorpoor, M., Idani, E., Khanjani, N., Goudarzi, G., Bahrampour, A., 2016. Relationship between air pollution, weather, traffic, and traffic-related mortality. Trauma monthly 21 (4).
- De Hoogh, K., Chen, J., Gulliver, J., Hoffmann, B., Hertel, O., Ketzel, M., Bauwelinck, M., van Donkelaar, A., Hvidtfeldt, U.A., Katsouyanni, K., 2018. Spatial PM2. 5, NO2, O3 and BC models for Western Europe–evaluation of spatiotemporal stability. Environ. Int. 120, 81–92.
- Faraji, M., Pourpak, Z., Naddafi, K., Nodehi, R.N., Nicknam, M.H., Shamsipour, M., Rezaei, S., Ghozikali, M.G., Ghanbarian, M., Mesdaghinia, A., 2018. Effects of airborne particulate matter (PM10) from dust storm and thermal inversion on global DNA methylation in human peripheral blood mononuclear cells (PBMCs) in vitro. Atmos. Environ. 195, 170–178.
- Faridi, S., Naddafi, K., Kashani, H., Nabizadeh, R., Alimohammadi, M., Momeniha, F., Faridi, S., Niazi, S., Zare, A., Gholampour, A., 2017. Bioaerosol exposure and circulating biomarkers in a panel of elderly subjects and healthy young adults. Sci. Total Environ. 593, 380–389.
- Faridi, S., Shamsipour, M., Krzyzanowski, M., Künzli, N., Amini, H., Azimi, F., Malkawi, M., Momeniha, F., Gholampour, A., Hassanvand, M.S., 2018. Long-term trends and health impact of PM 2.5 and O 3 in Tehran, Iran, 2006–2015. Environ. Int. 114, 37–49.
- Ghim, Y.S., Chang, Y.-S., Jung, K., 2015. Temporal and spatial variations in fine and coarse particles in Seoul, Korea. Aerosol Air Qual. Res. 15, 842–852.
- Guo, H., Wang, Y., Zhang, H., 2017. Characterization of criteria air pollutants in Beijing during 2014–2015. Environ. Res. 154, 334–344.
- Hassanvand, M.S., Naddafi, K., Faridi, S., Arhami, M., Nabizadeh, R., Sowlat, M.H., Pourpak, Z., Rastkari, N., Momeniha, F., Kashani, H., 2014. Indoor/outdoor relationships of PM10, PM2. 5, and PM1 mass concentrations and their water-soluble ions in a retirement home and a school dormitory. Atmos. Environ. 82, 375–382.
- Hassanvand, M.S., Naddafi, K., Faridi, S., Nabizadeh, R., Sowlat, M.H., Momeniha, F., Gholampour, A., Arhami, M., Kashani, H., Zare, A., 2015. Characterization of PAHs and metals in indoor/outdoor PM10/PM2. 5/PM1 in a retirement home and a school dormitory. Sci. Total Environ. 527, 100–110.
- Hassanvand, M.S., Naddafi, K., Kashani, H., Faridi, S., Kunzli, N., Nabizadeh, R., Momeniha, F., Gholampour, A., Arhami, M., Zare, A., 2017. Short-term effects of particle size fractions on circulating biomarkers of inflammation in a panel of elderly subjects and healthy young adults. Environ. Pollut. 223, 695–704.
- He, J., Gong, S., Yu, Y., Yu, L., Wu, L., Mao, H., Song, C., Zhao, S., Liu, H., Li, X., 2017. Air pollution characteristics and their relation to meteorological conditions during 2014–2015 in major Chinese cities. Environ. Pollut. 223, 484–496.
- Heger, M., Sarraf, M., 2018. Air Pollution in Tehran: Health Costs, Sources, and Policies. World Bank.
- Hoseini, M., Nabizadeh, R., Delgado-Saborit, J.M., Rafiee, A., Yaghmaeian, K., Parmy, S., Faridi, S., Hassanvand, M.S., Yunesian, M., Naddafi, K., 2018. Environmental and lifestyle factors affecting exposure to polycyclic aromatic hydrocarbons in the general population in a Middle Eastern area. Environ. Pollut. 240, 781–792.
- http://airnow.tehran.ir/home/dataarchive.aspx.
- http://traffic.tehran.ir/Default.aspx?tabid=152.
- Huang, R.-J., Zhang, Y., Bozzetti, C., Ho, K.-F., Cao, J.-J., Han, Y., Daellenbach, K.R., Slowik, J.G., Platt, S.M., Canonaco, F., 2014. High secondary aerosol contribution to particulate pollution during haze events in China. Nature 514 (7521), 218.
- Jhun, I., Coull, B.A., Zanobetti, A., Koutrakis, P., 2015. The impact of nitrogen oxides concentration decreases on ozone trends in the USA. Air Quality, Atmosphere & Health 8 (3), 283–292.
- Ji, G., Tian, L., Zhao, J., Yue, Y., Wang, Z., 2019. Detecting spatiotemporal dynamics of PM2. 5 emission data in China using DMSP-OLS nighttime stable light data. J. Clean. Prod. 209, 363–370.

- Jiang, L., Bai, L., 2018. Spatio-temporal characteristics of urban air pollutions and their causal relationships: evidence from Beijing and its neighboring cities. Sci. Rep. 8 (1), 1279.
- Jin, Q., Fang, X., Wen, B., Shan, A., 2017. Spatio-temporal variations of PM2. 5 emission in China from 2005 to 2014. Chemosphere 183, 429–436.
- Kim, E., Hopke, P.K., Pinto, J.P., Wilson, W.E., 2005. Spatial variability of fine particle mass, components, and source contributions during the regional air pollution study in St. Louis. Environmental Science & Technology 39 (11), 4172–4179.
- Krudysz, M.A., Froines, J.R., Fine, P.M., Sioutas, C., 2008. Intra-community spatial variation of size-fractionated PM mass, OC, EC, and trace elements in the Long Beach, CA area. Atmos. Environ. 42 (21), 5374–5389.
- Li, R., Cui, L., Li, J., Zhao, A., Fu, H., Wu, Y., Zhang, L., Kong, L., Chen, J., 2017. Spatial and temporal variation of particulate matter and gaseous pollutants in China during 2014–2016. Atmos. Environ. 161, 235–246.
- Mangia, C., Gianicolo, E.A., Bruni, A., Vigotti, M.A., Cervino, M., 2013. Spatial variability of air pollutants in the city of Taranto, Italy and its potential impact on exposure assessment. Environ. Monit. Assess. 185 (2), 1719–1735.
- Marlier, M.E., Jina, A.S., Kinney, P.L., DeFries, R.S., 2016. Extreme air pollution in global megacities. Current Climate Change Reports 2 (1), 15–27.
- Marzouni, M.B., Alizadeh, T., Banafsheh, M.R., Khorshiddoust, A.M., Ghozikali, M.G., Akbaripoor, S., Sharifi, R., Goudarzi, G., 2016. A comparison of health impacts assessment for PM10 during two successive years in the ambient air of Kermanshah, Iran. Atmospheric Pollution Research 7 (5), 768–774.
- Mohammadiha, A., Malakooti, H., Esfahanian, V., 2018. Development of reduction scenarios for criteria air pollutants emission in Tehran Traffic Sector, Iran. Sci. Total Environ. 622, 17–28.
- Naddafi, K., Hassanvand, M.S., Yunesian, M., Momeniha, F., Nabizadeh, R., Faridi, S., Gholampour, A., 2012. Health impact assessment of air pollution in megacity of Tehran, Iran. Iranian Journal of Environmental Health Science & Engineering 9 (1), 28.
- Pinto, J.P., Lefohn, A.S., Shadwick, D.S., 2004. Spatial variability of PM2. 5 in urban areas in the United States. J. Air Waste Manage. Assoc. 54 (4), 440–449.
- Saha, P.K., Zimmerman, N., Malings, C., Hauryliuk, A., Li, Z., Snell, L., Subramanian, R., Lipsky, E., Apte, J.S., Robinson, A.L., 2019. Quantifying high-resolution spatial variations and local source impacts of urban ultrafine particle concentrations. Sci. Total Environ. 655, 473–481.
- Schraufnagel, D.E., Balmes, J.R., Cowl, C.T., De Matteis, S., Jung, S.H., Mortimer, K., Perez-Padilla, R., Rice, M.B., Riojas-Rodriguez, H., Sood, A., Thurston, G.D., To, T., Vanker, A., Wuebbles, D.J., 2019. Air pollution and noncommunicable diseases: a review by the forum of international respiratory societies' environmental committee, part 2: air pollution and organ systems. Chest 155 (2), 417–426.
- Seifi, M., Niazi, S., Johnson, G., Nodehi, V., Yunesian, M., 2019. Exposure to ambient air pollution and risk of childhood cancers: a population-based study in Tehran, Iran. Sci. Total Environ. 646, 105–110.
- Shahbazi, H., Reyhanian, M., Hosseini, V., Afshin, H., 2016a. The relative contributions of mobile sources to air pollutant emissions in Tehran, Iran: an emission inventory approach. Emission Control Science and Technology 2 (1), 44–56.
- Shahbazi, H., Taghvaee, S., Hosseini, V., Afshin, H., 2016b. A GIS based emission inventory development for Tehran. Urban Clim. 17, 216–229.
- Shahbazi, H., Karimi, S., Hosseini, V., Yazgi, D., Torbatian, S., 2018. A novel regression imputation framework for Tehran air pollution monitoring network using outputs from WRF and CAMx models. Atmos. Environ. 187, 24–33.
- Song, C., Wu, L., Xie, Y., He, J., Chen, X., Wang, T., Lin, Y., Jin, T., Wang, A., Liu, Y., 2017. Air pollution in China: status and spatiotemporal variations. Environ. Pollut. 227, 334–347.

- Song, Y., Wang, X., Maher, B.A., Li, F., Xu, C., Liu, X., Sun, X., Zhang, Z., 2016. The spatialtemporal characteristics and health impacts of ambient fine particulate matter in China. J. Clean. Prod. 112, 1312–1318.
- Taghvaee, S., Sowlat, M.H., Hassanvand, M.S., Yunesian, M., Naddafi, K., Sioutas, C., 2018a. Source-specific lung cancer risk assessment of ambient PM2. 5-bound polycyclic aromatic hydrocarbons (PAHs) in central Tehran. Environ. Int. 120, 321–332.
- Taghvaee, S., Sowlat, M.H., Mousavi, A., Hassanvand, M.S., Yunesian, M., Naddafi, K., Sioutas, C., 2018b. Source apportionment of ambient PM2. 5 in two locations in central Tehran using the Positive Matrix Factorization (PMF) model. Sci. Total Environ. 628, 672–686.
- Taghvaee, S., Sowlat, M.H., Mousavi, A., Hassanvand, M.S., Yunesian, M., Naddafi, K., Sioutas, C., 2018c. Source apportionment of ambient PM 2.5 in two locations in central Tehran using the Positive Matrix Factorization (PMF) model. Sci. Total Environ. 628, 672–686.
- Tan, Y., Lipsky, E.M., Saleh, R., Robinson, A.L., Presto, A.A., 2014. Characterizing the spatial variation of air pollutants and the contributions of high emitting vehicles in Pittsburgh, PA. Environmental Science & Technology 48 (24), 14186–14194.
- Tiwari, S., Hopke, P.K., Pipal, A.S., Srivastava, A.K., Bisht, D.S., Tiwari, S., Singh, A.K., Soni, V.K., Attri, S.D., 2015. Intra-urban variability of particulate matter (PM2. 5 and PM10) and its relationship with optical properties of aerosols over Delhi, India. Atmos. Res. 166, 223–232.
- Wang, X., Qian, Z., Hong, H., Yang, Y., Xu, Y., Xu, X., Yao, Z., Zhang, L., Rolling, C.A., Schootman, M., 2018. Estimating the acute effects of fine and coarse particle pollution on stroke mortality of in six Chinese subtropical cities. Environ. Pollut. 239, 812–817.
- Wang, Y., Ying, Q., Hu, J., Zhang, H., 2014. Spatial and temporal variations of six criteria air pollutants in 31 provincial capital cities in China during 2013–2014. Environ. Int. 73, 413–422.
- Wilson, J.G., Kingham, S., Pearce, J., Sturman, A.P., 2005. A review of intraurban variations in particulate air pollution: implications for epidemiological research. Atmos. Environ. 39 (34), 6444–6462.
- Wongphatarakul, V., Friedlander, S., Pinto, J., 1998. A comparative study of PM2. 5 ambient aerosol chemical databases. Environmental Science & Technology 32 (24), 3926–3934.
- Xu, Y., Ying, Q., Hu, J., Gao, Y., Yang, Y., Wang, D., Zhang, H., 2018. Spatial and temporal variations in criteria air pollutants in three typical terrain regions in Shaanxi, China, during 2015. Air Quality, Atmosphere & Health 11 (1), 95–109.
- Yousefian, F., Mahvi, A.H., Yunesian, M., Hassanvand, M.S., Kashani, H., Amini, H., 2018. Long-term exposure to ambient air pollution and autism spectrum disorder in children: a case-control study in Tehran, Iran. Sci. Total Environ. 643, 1216–1222.
- Zhang, N.-N., Ma, F., Qin, C.-B., Li, Y.-F., 2018. Spatiotemporal trends in PM2. 5 levels from 2013 to 2017 and regional demarcations for joint prevention and control of atmospheric pollution in China. Chemosphere 210, 1176–1184.
- Zhang, N.-N., Ma, F., Guan, Y., Li, Y.-F., 2019. Spatial assessment of air resources in China from 2013 to 2017. Sci. Total Environ. 658, 294–304.
- Zhang, Y.-L, Cao, F., 2015. Fine particulate matter (PM 2.5) in China at a city level. Sci. Rep. 5, 14884.
- Zhao, D., Chen, H., Li, X., Ma, X., 2018. Air pollution and its influential factors in China's hot spots. J. Clean. Prod. 185, 619–627.
- Zhao, S., Yu, Y., Yin, D., He, J., Liu, N., Qu, J., Xiao, J., 2016. Annual and diurnal variations of gaseous and particulate pollutants in 31 provincial capital cities based on in situ air quality monitoring data from China National Environmental Monitoring Center. Environ. Int. 86, 92–106.
- van Zoest, V., Stein, A., Hoek, G., 2018. Outlier detection in urban air quality sensor networks. Water Air Soil Pollut. 229 (4), 111.